

Ecohydrology & Hydrobiology

The influence of locks on zooplankton in canals (Bydgoszcz canal and Noteć canal, Poland)

--Manuscript Draft--

Manuscript Number:	
Article Type:	Original article
Keywords:	Rotifers; Crustaceans; Bydgoszcz Canal; water flow; environmental parameters
Corresponding Author:	Nikola Kolarova Kazimierz Wielki University in Bydgoszcz Faculty of Biological Sciences Bydgoszcz, Kujawsko-Pomorskie POLAND
First Author:	Nikola Kolarova
Order of Authors:	Nikola Kolarova Pawel Napiórkowski
Abstract:	<p>Artificial waterways are characterized by series of dams and locks that slow the flow of water. Such conditions can influence the structure of zooplankton communities. The aim of our research was to assess the impact of environmental and hydrological conditions on the zooplankton of sites upstream and downstream of canal locks. The study was carried out in 2021 and 2022. Water samples were collected monthly from the Bydgoszcz Canal and the Noteć Canal (Poland) during the growing season. We evaluated how water flow velocity and selected environmental parameters (i.e., water temperature, Secchi disk visibility, conductivity, oxygen concentration, saturation, pH and chl-a concentrations) influence the zooplankton diversity (T) density (N) and biomass (B). The results of our study showed that density and biomass of rotifers changed approximately proportionally to changes in chlorophyll at sites upstream of locks. Zooplankton diversity also increased upstream of locks. A habitat of low flow velocity and characterized by macrophytes favored high diversity and density of zooplankton. The growth of crustaceans (density and biomass) as well as total zooplankton biomass was affected by water temperature at sites downstream of the lock. Sites 2 and 3 share a pattern of abundance and biomass. Site 1 differs probably due to a one-off increase in the number of crustaceans that probably originated within inside the lock. This was likely the result of internal loading (organic matter and nutrients) after the re-suspension of bottom sediments due to increased water movement inside the lock. Water enriched in suspension drained below the sluice.</p>
Suggested Reviewers:	Agnieszka Pocięcha pocięcha@iop.krakow.pl Andrzej Gołdyn rgold@amu.edu.pl Natalia Kuczynska-Kippen nkippen@amu.edu.pl Griselda Chaparro grichaparro@gmail.com Robert Czerniawski Robert.Czerniawski@usz.edu.pl



KAZIMIERZ WIELKI UNIVERSITY
IN BYDGOSZCZ

Faculty of Biological Sciences

Department of Hydrobiology
Chodkiewicz 30f Str., 85-064 Bydgoszcz, Poland
tel. +48(52) 34 19 171



Bydgoszcz, 2023-06-27

Dear Editors,

This cover letter accompanies our paper **The influence of locks on zooplankton in canals (Bydgoszcz canal and Noteć canal, Poland)**

authored by *Nikola Kolarova and Paweł Napiórkowski*.

Article deals with the impact of environmental and hydrological conditions on the zooplankton of sites upstream and downstream of canal locks. The study was carried out in 2021 and 2022. Water samples were collected monthly from the Bydgoszcz Canal and the Noteć Canal (Poland) during the growing season. We chose three sites where samples were taken upstream and downstream of the locks. We evaluated how water flow velocity and selected environmental parameters (i.e., water temperature, Secchi disk visibility, conductivity, oxygen concentration, saturation, pH and chl-a concentrations) influence the zooplankton diversity (T) density (N) and biomass (B). The results of our study showed that density and biomass of rotifers changed approximately proportionally to changes in chlorophyll at sites upstream of locks. Zooplankton diversity also increased upstream of locks. A habitat of low flow velocity and characterized by macrophytes favored high diversity and density of zooplankton. The growth of crustaceans (density and biomass) as well as total zooplankton biomass was affected by water temperature at sites downstream of the locks. We found that hydrotechnical constructions on canals have a significant impact on the structure of zooplankton. We decided to do this study because few literature items were devoted to zooplankton of artificial waterways.

We believe that our paper is relevant to a broad international audience and well suited for the readers of **Ecohydrology and Hydrobiology**.

We hope that you will consider this manuscript for review and we look forward to your response.

Yours sincerely,

Nikola Koralova
Paweł Napiórkowski

Msc. Nikola Kolarova www.ukw.edu.pl
Dept. of Hydrobiology
e-mail: nikol77@student.ukw.edu.pl
tel. 0048 523419 171

**The influence of locks on zooplankton in canals (Bydgoszcz canal and Noteć canal,
Poland)**

Nikola Kolarova*, Paweł Napiórkowski

Department of Hydrobiology, Faculty of Biological Sciences, Kazimierz Wielki University, Ossolińskich 12 Street, 85-093 Bydgoszcz, Poland, pnapiork@ukw.edu.pl, ORCID: 0000-0003-1987-9468

* Correspondence: Nikola Kolarova, Department of Hydrobiology, Faculty of Biological Sciences, Kazimierz Wielki University, Ossolińskich 12 Street, 85-093 Bydgoszcz, Poland

e-mail address: nikol77@student.ukw.edu.pl, ORCID: 0000-0002-8847-1063

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

**The influence of locks on zooplankton in canals (Bydgoszcz canal and Noteć canal,
Poland)**

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

The influence of locks on zooplankton in canals (Bydgoszcz canal and Noteć canal, Poland)

Abstract

Artificial waterways are characterized by series of dams and locks that slow the flow of water. Such conditions can influence the structure of zooplankton communities.

The aim of our research was to assess the impact of environmental and hydrological conditions on the zooplankton of sites upstream and downstream of canal locks.

The study was carried out in 2021 and 2022. Water samples were collected monthly from the Bydgoszcz Canal and the Noteć Canal (Poland) during the growing season. We evaluated how water flow velocity and selected environmental parameters (i.e., water temperature, Secchi disk visibility, conductivity, oxygen concentration, saturation, pH and chl-a concentrations) influence the zooplankton diversity (T) density (N) and biomass (B).

The results of our study showed that density and biomass of rotifers changed approximately proportionally to changes in chlorophyll at sites upstream of locks. Zooplankton diversity also increased upstream of locks. A habitat of low flow velocity and characterized by macrophytes favored high diversity and density of zooplankton. The growth of crustaceans (density and biomass) as well as total zooplankton biomass was affected by water temperature at sites downstream of the lock.

Sites 2 and 3 share a pattern of abundance and biomass. Site 1 differs probably due to a one-off increase in the number of crustaceans that probably originated within inside the lock. This was likely the result of internal loading (organic matter and nutrients) after the re-suspension of bottom sediments due to increased water movement inside the lock. Water enriched in suspension drained below the sluice.

Keywords: Rotifers, Crustaceans, Bydgoszcz Canal, water flow, environmental parameters.

Introduction

Artificial canals are important waterways connecting natural rivers into a large-scale inland water system (Gorączko, 2015). Such an artificial waterway is divided into sections that differ in water table height by means of locks, lifts or slipways. Many canals have a series of dams and locks that slow the flow of water, creating small reservoirs (Bydgoszcz canal). They often have an important navigational role in providing navigability past weirs and dams by maintaining appropriate water levels and depths while also allowing boats to ascend and descend between sections of differing elevations (Segovia et al., 2019).

Locks are inherent structures of canals whose main function is to raise water levels and enable travel in both directions, i.e., transporting boats across a water level differential – from lower to higher and *vice versa*. In combination with the construction of artificial canals and regulation of rivers, the building of locks allowed inland transport water connections to be developed (Muszyńska-Jeleszyńska & Marciniak, 2016).

Canals vary in hydrological regime, environmental conditions and organism communities. They are useful corridors for the spread of aquatic species (Kim & Mandrak, 2016; Pagnucco et al., 2015). These waterways provide direct (dispersal) or indirect (shipping) transportation routes for non-indigenous species, including of planktonic organisms (Dexter et al., 2020). Some studies report that large rivers connected with canals differ in terms of hydrological and environmental characteristics, what influence the diversity of aquatic biota such as zooplankton species (Appel et al., 2020; Kelly et al., 2013; Ball et al., 2018), phytoplankton species (Dembowska, 2021; Kelly & Hassall, 2018), macrophytes (Tarkowska-Kukuryk & Grzywna, 2022; Dorotovičová, 2013), benthic invertebrates (McCabe et al., 1998; Weber, 2017) and fish (Chester & Robson, 2013; Onikura, 2015; Roberts & Rahel, 2008). The water flow velocity near locks determines the reproductive success of aquatic organisms and consequently influences growing potential.

Zooplankton organisms can affect biodiversity in canals (Furst et al., 2014). Locks reduce water flow in canals and thus increase the number of zooplankton species. Artificial waterways within large river systems provide shelter from the high flow conditions of an open river, allowing large-bodied crustacean zooplankton to flourish (Dickerson et al., 2010). Locks may serve as retention areas and sustain crustacean species in the canals. Some fast-reproducing, small-bodied groups such as rotifers can be found in high densities in higher-flow conditions in canals (Havel et al., 2009).

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Freshwater zooplankton have an essential position in aquatic food webs as primary and secondary consumers (Majagi et al., 2019). They take part in nutrient cycling, regulate the biomass of phytoplankton (Gianuca et al., 2016) and provide food for fish (Ning et al., 2010), especially for larval stages and fish fry (Medeiros & Arthington, 2008). The zooplankton community is essential to ensuring a healthy ecosystem. Each species differently affects ecosystem functioning (Venkatraju et al., 2010; Manickam et al., 2018; Jeelani et al., 2008; Symons & Arnott, 2013).

Hydrological conditions are an important factor shaping the structure and functioning of zooplankton community in rivers and other waterways (Napiórkowski et al., 2019; Zhao et al., 2018; Czerniawski & Sługocki, 2017). Canals seem to be transitional between flowing and stagnant waters. Therefore, hydrological conditions may vary depending on the nature of the water flow (velocity, direction, average and seasonal variation) in the canal, but also depending on the distance between the locks and the way the locks work.

The aim of the study was to compare zooplankton species compositions between sites upstream and downstream of locks in two artificial waterways. We also compare the environmental conditions upstream and downstream of the locks and their impact on the zooplankton communities.

We assumed that zooplankton communities would differ in diversity, density and biomass between the two sides of the lock based on the differences in hydrological and environmental conditions.

We hypothesized that the canal waters upstream of the lock will stagnate, thereby creating better conditions for the development of zooplankton. Conversely, we hypothesized that the water flowing down from the lock would temporarily increase the water flow velocity, resulting in a decrease in zooplankton diversity and density.

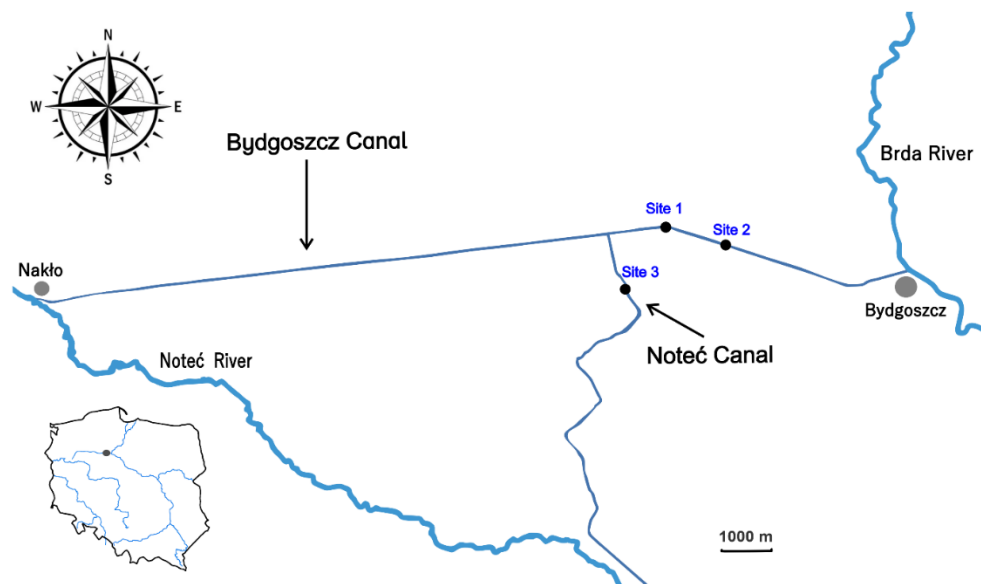
Materials and methods

The study was carried out during the growing season from April to September in 2021 and 2022 in the Bydgoszcz Canal (part of which is located in the industrial zone of Bydgoszcz city) and the Noteć Canal (located in the agricultural rural zone near the town of Nakło). The study was conducted in close proximity to (downstream and upstream of) locks on both canals. Two locks are located on the Bydgoszcz Canal (Osowa Góra and Prądy) and one is located in the Noteć Canal (Łochowo). The Bydgoszcz Canal is located in north-west of Poland between

1 Bydgoszcz city and Nakło town. The canal was built at the end of the 18th century. It is an
2 important part of the E70 international waterway linking the water systems of Western, Central
3 and Eastern Europe. This artificial watercourse connects the two largest rivers in Poland – the
4 Vistula and the Oder – through their tributaries. The total length of the canal is 24.7 km, of
5 which 15.7 km is located in the Noteć catchment (a tributary of the Odra River) and 9.0 km in
6 the Brda catchment (a tributary of the Vistula River).
7
8
9

10
11 The Osowa Góra and Prądy locks were initially built at the end of 18th century.
12 However, their final shape was determined during reconstruction works at the beginning of the
13 20th century. The locks water elevations are, respectively, 3.55 m up to 3.82 m, the total height
14 of walls is 7 m and width is about 29 m. The average water depth was 1.5 m at the sites upstream
15 of the locks and 1.7 m downstream. The water flow velocity was ~0.07 m/s at sites upstream
16 of the lock and ~0.004 m/s downstream.
17
18
19
20
21

22 The Noteć Canal consists of two parts (one referred to as simply the “Noteć Canal” and
23 the other as the “Upper Noteć Canal”). This waterway covers the course of the Noteć River.
24 The total length of the canal is 25 km. It is characterized by low water discharge and a strong
25 impact of anthropogenic contaminants due to human activities, including agriculture. The
26 Łochowo lock on the Upper Noteć Canal is ~15 m wide. The average depth was 1.4 m at the
27 site upstream of the lock and 1.6 m downstream. The water flow velocity was ~0.14 m/s at the
28 site upstream of the lock and ~0.3 m/s downstream.
29
30
31
32
33
34



54 **Figure 1.** Map of investigated area. Bydgoszcz Canal: site 1 – Osowa Góra; site 2 – Prądy; site
55 3 – Noteć Canal
56

57
58 Water samples were collected at three sampling sites: in the Bydgoszcz Canal: (site 1) Osowa
59 Góra upstream and downstream of the lock 53°08'48.9"N 17°52'49.2"E; (site 2) Prądy
60
61
62
63
64
65

1 upstream and downstream of the lock 53°08'38.6"N 17°53'37.8"E and (site 3) in the Noteć
2 Canal- Lochowo upstream and downstream of the lock 53°07'56.5"N 17°51'18.1"E. Samples
3 were collected from different depths with a 1-L Patalas bucket once a month, for a total of 36
4 samples. To obtain one qualitative and quantitative sample of zooplankton, 20 L of water was
5 filtered through a plankton net of mesh size 25 µm. All samples were preserved with Lugol's
6 iodine solution (Harris et al., 2000; Wallence et al., 1993). The zooplankton density was
7 determined under a microscope using a Sedgewick Rafter-type chamber (1 mL) by a
8 methodology after McCauley (1984) and counted per 1 L of water. The rotifers' wet weight
9 was calculated using the formula according to Radwan (2004). The zooplankton species were
10 identified using an Olympus BX 43 light microscope and an Olympus LC 30 soft imaging
11 camera at 10× magnification according to commonly available keys (Wallace et al., 1993;
12 Einsle, 1996; Radwan et al., 2004; Błędzki & Rybak, 2016). Concurrently with zooplankton
13 sampling, the selected environmental parameters of water were measured: water temperature
14 (WT, °C), Secchi disk visibility (SD, m), conductivity (EC, µS cm⁻¹), oxygen concentration
15 (DO, mg/L), saturation (%), chlorophyll (chl-a, µg/L) and pH (Table 1). In the laboratory,
16 biological material was poured into a glass 25-ml cuvette and analyzed using a spectral ALA
17 fluorimeter (AlgaeLabAnalyser, BBE Germany). One measurement was an arithmetic mean of
18 three "fast analyses". In this way, we collected data about total Chl-a concentration (TChl-a,
19 µg/L) and its concentration in four taxonomic groups: the Chlorophyta, Bacillariophyta,
20 Cyanobacteria, and Cryptophyta (µg/L). For proper calculation of TChl-a, we corrected for
21 yellow substances using the chromophoric dissolved organic matter correction. The whole
22 procedure was performed within 72 h from *in-situ* sample collection. For a detailed description,
23 see Nguyen et al. (2015). Multimeter WTW Multi 3430SET F Xylem Analytics field probes
24 (Weilheim, Germany) were used for measurements. During the sampling period, the surface
25 water flow was measured using the electromagnetic hydrometric mill (Model 801).
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44

45 The hierarchical clustering analysis and redundancy analyses (RDA) were used to
46 determine the environmental variables responsible for variations in the zooplankton taxonomic
47 composition, density and biomass during the growing seasons at the study sites.
48
49
50

51 The explanatory response variables used in the analyses were: WT, SD, pH, DO, EC,
52 Chl-a and number of zooplankton species (total number of zooplankton species, number of
53 rotifer and crustacean species), zooplankton density (total zooplankton density, rotifer and
54 crustacean density), and zooplankton biomass (total zooplankton biomass, rotifer and
55 crustacean density).
56
57
58
59
60
61
62
63
64
65

1 crustacean biomass). For statistical analysis were used all data variability (environmental and
2 biological) including all investigated months.

3
4 Pearson simple coefficient was tested to analyze statistically significant correlations
5 between environmental and biological parameters by Past 4.03 software (Hammer et al., 2001).
6 The hierarchical clustering analysis was performed by grouping the sites downstream and
7 upstream of the lock depending on their similarity in terms of environmental and biological
8 data in Past 4.03 software (Hammer et al., 2001). The pairwise distances between
9 environmental parameters were measured by Euclidean similarity index and between
10 biological parameters by the Bray–Curtis similarity index. The dataset was $\log(x+1)$
11 transformed to minimize differences in variables.
12
13
14
15
16
17

18 Redundancy analysis (RDA) was performed using the constrained linear method to
19 determine the relationships between selected environmental parameters and biological data
20 (Rao, 1964; van den Wollenberg, 1977). The dataset was $\log(x+1)$ transformed to eliminate
21 the effect of outliers on the results. The data used in RDA analysis were taken from the
22 Bydgoszcz Canal and the Noteć Canal at sites downstream and upstream of each lock. The
23 presented results were processed statistically by using Canoco 5.0 software (ter Braak &
24 Šmilauer, 2002).
25
26
27
28
29
30

31 32 **Results**

33 34 35 36 *Physico-chemical parameters*

37
38
39
40 The average water temperature, concentration of dissolved oxygen, saturation and of pH value
41 at sites in the Bydgoszcz Canal and in the Noteć Canal were similar upstream of the locks to
42 the average values recorded at sites downstream. The average Secchi disk visibility differed at
43 Site 3. At this site, 1.5 m was recorded upstream of the lock and slightly less (1.4 m)
44 downstream. The average conductivity was higher at all upstream sites. The greatest
45 differences were found at Site 1 upstream. The highest value of this parameter was $2684 \mu\text{S cm}^{-1}$
46 while the lowest value was $595 \mu\text{S cm}^{-1}$. The average chlorophyll concentration varied
47 most widely at Site 1. The average concentration ranged from $21.2 \mu\text{g/L}$ upstream to $31.2 \mu\text{g/L}$
48 downstream. The highest concentration of chlorophyll ($103.3 \mu\text{g/L}$) was recorded at Site 2
49 upstream, and the lowest concentration ($1.6 \mu\text{g/L}$) was recorded at Site 3 downstream (Table
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
66
67
68
69
70
71
72
73
74
75
76
77
78
79
80
81
82
83
84
85
86
87
88
89
90
91
92
93
94
95
96
97
98
99
100
101
102
103
104
105
106
107
108
109
110
111
112
113
114
115
116
117
118
119
120
121
122
123
124
125
126
127
128
129
130
131
132
133
134
135
136
137
138
139
140
141
142
143
144
145
146
147
148
149
150
151
152
153
154
155
156
157
158
159
160
161
162
163
164
165
166
167
168
169
170
171
172
173
174
175
176
177
178
179
180
181
182
183
184
185
186
187
188
189
190
191
192
193
194
195
196
197
198
199
200
201
202
203
204
205
206
207
208
209
210
211
212
213
214
215
216
217
218
219
220
221
222
223
224
225
226
227
228
229
230
231
232
233
234
235
236
237
238
239
240
241
242
243
244
245
246
247
248
249
250
251
252
253
254
255
256
257
258
259
260
261
262
263
264
265
266
267
268
269
270
271
272
273
274
275
276
277
278
279
280
281
282
283
284
285
286
287
288
289
290
291
292
293
294
295
296
297
298
299
300
301
302
303
304
305
306
307
308
309
310
311
312
313
314
315
316
317
318
319
320
321
322
323
324
325
326
327
328
329
330
331
332
333
334
335
336
337
338
339
340
341
342
343
344
345
346
347
348
349
350
351
352
353
354
355
356
357
358
359
360
361
362
363
364
365
366
367
368
369
370
371
372
373
374
375
376
377
378
379
380
381
382
383
384
385
386
387
388
389
390
391
392
393
394
395
396
397
398
399
400
401
402
403
404
405
406
407
408
409
410
411
412
413
414
415
416
417
418
419
420
421
422
423
424
425
426
427
428
429
430
431
432
433
434
435
436
437
438
439
440
441
442
443
444
445
446
447
448
449
450
451
452
453
454
455
456
457
458
459
460
461
462
463
464
465
466
467
468
469
470
471
472
473
474
475
476
477
478
479
480
481
482
483
484
485
486
487
488
489
490
491
492
493
494
495
496
497
498
499
500
501
502
503
504
505
506
507
508
509
510
511
512
513
514
515
516
517
518
519
520
521
522
523
524
525
526
527
528
529
530
531
532
533
534
535
536
537
538
539
540
541
542
543
544
545
546
547
548
549
550
551
552
553
554
555
556
557
558
559
560
561
562
563
564
565
566
567
568
569
570
571
572
573
574
575
576
577
578
579
580
581
582
583
584
585
586
587
588
589
590
591
592
593
594
595
596
597
598
599
600
601
602
603
604
605
606
607
608
609
610
611
612
613
614
615
616
617
618
619
620
621
622
623
624
625
626
627
628
629
630
631
632
633
634
635
636
637
638
639
640
641
642
643
644
645
646
647
648
649
650
651
652
653
654
655
656
657
658
659
660
661
662
663
664
665
666
667
668
669
670
671
672
673
674
675
676
677
678
679
680
681
682
683
684
685
686
687
688
689
690
691
692
693
694
695
696
697
698
699
700
701
702
703
704
705
706
707
708
709
710
711
712
713
714
715
716
717
718
719
720
721
722
723
724
725
726
727
728
729
730
731
732
733
734
735
736
737
738
739
740
741
742
743
744
745
746
747
748
749
750
751
752
753
754
755
756
757
758
759
760
761
762
763
764
765
766
767
768
769
770
771
772
773
774
775
776
777
778
779
780
781
782
783
784
785
786
787
788
789
790
791
792
793
794
795
796
797
798
799
800
801
802
803
804
805
806
807
808
809
810
811
812
813
814
815
816
817
818
819
820
821
822
823
824
825
826
827
828
829
830
831
832
833
834
835
836
837
838
839
840
841
842
843
844
845
846
847
848
849
850
851
852
853
854
855
856
857
858
859
860
861
862
863
864
865
866
867
868
869
870
871
872
873
874
875
876
877
878
879
880
881
882
883
884
885
886
887
888
889
890
891
892
893
894
895
896
897
898
899
900
901
902
903
904
905
906
907
908
909
910
911
912
913
914
915
916
917
918
919
920
921
922
923
924
925
926
927
928
929
930
931
932
933
934
935
936
937
938
939
940
941
942
943
944
945
946
947
948
949
950
951
952
953
954
955
956
957
958
959
960
961
962
963
964
965
966
967
968
969
970
971
972
973
974
975
976
977
978
979
980
981
982
983
984
985
986
987
988
989
990
991
992
993
994
995
996
997
998
999
1000

Table 1. Mean values of environmental parameters in Bydgoszcz Canal and Noteć Canal, upstream and downstream of the lock. Water temperature (WT, °C), Secchi disk visibility (SD, m), conductivity (EC, $\mu\text{S cm}^{-1}$), oxygen concentration (DO, mg/L), saturation (%), pH, chlorophyll (chl-a, $\mu\text{g/L}$)

	Bydgoszcz Canal				Noteć Canal	
	Site 1 A		Site 2 A		Site 3 A	
	mean	range	Mean	range	mean	range
WT (°C)	18.1	(10.1–25.2)	18.2	(10.8–25.9)	17.4	(8.9–25.1)
SD (m)	1.6	(0.6–2.1)	1.5	(0.4–2.3)	1.5	(0.6–2.0)
EC ($\mu\text{S cm}^{-1}$)	1471	(595–2684)	1324	(595–2030)	1349	(593–2137)
DO (mg/L)	9.4	(3.5–16.5)	10.3	(5.9–17.0)	8.6	(5.4–13.3)
SAT (%)	107	(63–179)	119	(86–181)	83	(67–107)
pH	7.7	(6.8–8.3)	7.7	(6.8–8.5)	7.7	(6.4–8.4)
Chl-a ($\mu\text{g/L}$)	21.2	(2.2–81.7)	34.9	(7.5–103.3)	22.9	(1.9–87.7)

	Bydgoszcz Canal				Noteć Canal	
	Site 1 B		Site 2 B		Site 3 B	
	mean	range	Mean	range	mean	range
WT (°C)	17.5	(8.6–26.2)	17.3	(8.9–25.8)	16.0	(8.0–24.3)
SD (m)	1.6	(0.6–2.2)	1.5	(0.4–2.2)	1.4	(0.6–1.9)
EC ($\mu\text{S cm}^{-1}$)	938	(598–1812)	925	(592–1805)	949	(594–1840)
DO (mg/L)	10.3	(7.9–15.8)	10.3	(7.6–14.0)	8.8	(6.5–12.0)
SAT (%)	109	(81–170)	108	(84–148)	87	(67–100)
pH	7.6	(6.8–8.4)	7.5	(6.7–8.5)	7.6	(6.5–8.6)
Chl-a ($\mu\text{g/L}$)	31.2	(2.8–100.2)	36.6	(8.0–102.7)	19.5	(1.6–84.2)

(sites A – upstream of the locks, sites B – downstream of the locks)

Zooplankton diversity, density and biomass

During the study, a total of 119 zooplankton species were identified. The highest richness was of rotifers, with 93 species (i.e., 78% of all species) followed by crustaceans, with 26 species (i.e., 22% of all species). At sites upstream, 103 zooplankton species were recorded, comprising 82 rotifer species (i.e., 80% of all species) and 21 crustacean species (i.e., 20% of all species) alongside nauplii and copepodites (larval forms of copepods). At sites downstream, 89 zooplankton species were recorded, comprising 71 rotifer species (i.e., 80% of all species) and 18 crustacean species (i.e., 20% of all species) accounted with larval forms of copepods. The highest total number of species was recorded at Site 1 upstream (82, comprising 66 rotifer species and 16 crustacean species). The lowest number of species was recorded at Site 2 and Site 3 downstream (61, comprising 51 rotifer species and 10 crustacean species) (Table 2). The

complete list of zooplankton species identified during the study period is provide in the supplementary materials. The highest number of species in a single sample was observed at Site 1 upstream (21 species) while downstream the lock (17 species). The lowest number of species was recorded at Site 2 downstream (16 species) while upstream (20 species) (Table 3).

Table 2. Total number of species (diversity) and dominants (density) in the zooplankton community during 2021 and 2022 growing seasons in Bydgoszcz Canal and Noteć Canal sites

	Bydgoszcz Canal				Noteć Canal	
	Site 1 A	Site 1 B	Site 2 A	Site 2 B	Site 3 A	Site 3 B
Rotifers	66	51	65	51	53	51
Crustaceans	16	11	11	10	15	10
Total	82	62	76	61	68	61
Dominant species and percent of domination	<i>Keratella cochlearis</i> * 64%	<i>Keratella cochlearis</i> * 40%	<i>Keratella cochlearis</i> * 50%	<i>Keratella cochlearis</i> * 66%	<i>Keratella cochlearis</i> * 71%	<i>Keratella cochlearis</i> * 72%
	<i>Keratella quadrata</i> * 9%	<i>Keratella quadrata</i> * 5%	<i>Keratella quadrata</i> * 6%	<i>Keratella quadrata</i> * 5%	<i>Keratella quadrata</i> * 6%	<i>Keratella quadrata</i> * 12%
	<i>Euchlanis dilatata</i> * 15%	<i>Polyarthra dolichoptera</i> * 2%	<i>Brachionus calyciflorus</i> * 3%	<i>Keratella tecta</i> * 4%	<i>Keratella tecta</i> * 8%	<i>Keratella tecta</i> * 8%
	<i>Chydorus sphaericus</i> ** 7%	<i>Bosmina longirostris</i> ** 50%	<i>Bosmina longirostris</i> ** 33%	<i>Bosmina longirostris</i> ** 20%	<i>Bosmina longirostris</i> ** 5%	<i>Chydorus sphaericus</i> ** 2%
	nauplius** 5%	nauplius** 3%	<i>Ceriodaphnia pulchella</i> ** 8%	nauplius** 5%	<i>Chydorus sphaericus</i> ** 10%	nauplius** 6%

(*rotifer, **crustacean, sites A – upstream of locks, sites B – downstream of locks)

The average zooplankton density was highest at Site 1 downstream of the lock (456 ind/L) and was two times higher than upstream (227 ind/L). The lowest average zooplankton density was at Site 3 downstream (172 ind/L) and was almost two times lower than upstream (261 ind/L). Average rotifer density was highest at Site 1 downstream (226 ind/L) and was higher than upstream (195 ind/L). The lowest average rotifer density was at Site 3 downstream (158 ind/L) and was lower than upstream (216 ind/L). Average crustacean density was highest at Site 1 downstream (230 ind/L) and was seven times higher than upstream (32 ind/L). The lowest

1
2 average crustacean density was found at Site 3 downstream (14 ind/L) and was three times
3 lower than upstream (45 ind/L) (Table 3).

4 During the study rotifers, were the most dominant (in terms of diversity and abundance)
5 in all collected samples. *Keratella cochlearis* (from 40% to 72%), *Keratella quadrata* (from
6 5% to 12%), *Keratella tecta* (~4%) prevailed at all sites. Among crustacean species, the most
7 dominant were *Bosmina longirostris* (from 2% to 50%), *Chydorus sphaericus* (from 2% to
8 10%) and nauplii (from 3% to 6%) (Table 2).

9 The average zooplankton biomass was highest at Site 1 downstream (4.321 mg/L) and
10 was five times higher than upstream (0.825 mg/L). The lowest average zooplankton biomass
11 was at Site 3 downstream (0.217 mg/L) and was five times lower than upstream (1.024 mg/L).
12 Average rotifer biomass was highest at Site 2 upstream (0.101 mg/L) and lowest at the same
13 site downstream (0.053 mg/L). The lowest average rotifer biomass was also recorded at Site 3
14 downstream (0.053 mg/L) which was slightly lower than upstream (0.056 mg/L). Average
15 crustacean biomass was the highest at Site 1 upstream (4.250 mg/L) and was almost six times
16 higher than downstream (0.728 mg/L). The lowest crustacean biomass was at Site 3 downstream
17 (0.164 ind/L) and was eight times lower than upstream (0.968 ind/L) (Table 3). In terms of
18 biomass, crustacean was the most dominant in almost all collected samples. *B. longirostris*
19 (61%), *Ceriodaphnia pulchella* (13%) and *Ch. Sphaericus* (8%) prevailed at all sites. Among
20 rotifer species *K. cochlearis* (28%), *Euchlanis dilatata* (19%), *K. quadrata* (17%) were the
21 most dominant.

22 At upstream sites, the α -diversity index ($H'=2.23\pm 0.20$) was at Site 1 and lowest
23 ($H'=1.84\pm 0.19$) at Site 2. At sites downstream α -diversity index ($H'=2.21\pm 0.19$) was highest at
24 Site 3 and lowest ($H'=1.92\pm 0.10$) at Site 1 and Site 2.

25 At upstream sites, the evenness index ($J'=0.52\pm 0.06$) was highest at Site 1 and lowest
26 ($J'=0.40\pm 0.06$) at Site 2. At downstream sites, the evenness index ($J'=0.54\pm 0.03$) was highest at
27 Site 3 and lowest ($H'=0.49 \pm 0.02$) at Site 2 (Table 3).

28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55 **Table 3.** Mean values of zooplankton species, density (ind/L) and biomass (mg/L) in
56 Bydgoszcz Canal and Noteć Canal, upstream and downstream of the lock. Tax total – number
57 of zooplankton species; Tax Rot – number of rotifer species; Tax Crust – number of crustacean
58
59
60
61
62
63
64
65

species; N total – density of zooplankton; N Rot – density of rotifers; N Crust – density of crustaceans; B total – biomass of zooplankton; B Rot – biomass of rotifers; B Crust – biomass of crustaceans. Shannon–Weaver α -diversity index (H' index); Pielou's evenness index (J' index)

	Bydgoszcz Canal				Noteć Canal	
	Site 1 A		Site 2 A		Site 3 A	
	mean	range	mean	range	mean	range
Tax total	21	(13–38)	20	(11–30)	19	(13–28)
Tax Rot	17	(7–30)	16	(8–23)	15	(8–24)
Tax Crust	4	(2–8)	4	(2–7)	4	(2–9)
N total	227	(12–1488)	333	(16–1462)	261	(17–1378)
N Rot	195	(5–1402)	211	(15–1409)	216	(8–1290)
N Crust	32	(1–86)	122	(1–876)	45	(2–223)
B total	0.825	(0.021–3.117)	2.873	(0.014–23.037)	1.024	(0.031–2.697)
B Rot	0.097	(0.002–0.491)	0.101	(0.004–0.544)	0.056	(0.002–0.231)
B Crust	0.728	(0.011–2.973)	2.772	(0.008–23.001)	0.968	(0.028–2.626)
H' index	2.23	(1.01–2.89)	1.84	(0.38–2.62)	2.02	(0.83–2.76)
J' index	0.52	(0.11–0.82)	0.40	(0.11–0.72)	0.47	(0.12–0.82)

	Bydgoszcz Canal				Noteć Canal	
	Site 1 B		Site 2 B		Site 3 B	
	mean	range	mean	range	mean	range
Tax total	17	(5–27)	16	(13–20)	19	(15–24)
Tax Rot	13	(1–20)	12	(7–16)	15	(12–21)
Tax Crust	4	(2–9)	4	(2–7)	4	(2–5)
N total	456	(12–1980)	258	(18–1507)	172	(28–962)
N Rot	226	(4–1912)	193	(14–1445)	158	(15–925)
N Crust	230	(2–1941)	65	(3–293)	14	(5–37)
B total	4.321	(0.026–36.333)	1.304	(0.023–5.388)	0.217	(0.042–0.664)
B Rot	0.068	(0.001–0.503)	0.053	(0.004–0.320)	0.053	(0.004–0.199)
B Crust	4.250	(0.014–36.320)	1.251	(0.019–5.381)	0.164	(0.029–0.488)
H' index	1.92	(0.07–2.75)	1.92	(0.71–2.61)	2.21	(0.93–2.69)
J' index	0.50	(0.11–0.83)	0.49	(0.12–0.77)	0.54	(0.11–0.77)

(sites A – upstream of the locks, sites B – downstream of the locks)

Table 4. Pearson's simple correlation r coefficient for environmental variables and zooplankton species composition

	Tax total	Tax Rot	Tax Crust	N total	N Rot	N Crust	B total	B Rot	B Crust
Sites A									
WT	0.28	0.18	0.28	-0.14	-0.25	0.11	0.02	-0.20	0.07
SD	-0.22	-0.24	-0.12	<u>-0.72</u>	<u>-0.83</u>	-0.36	-0.28	<u>-0.84</u>	-0.23
EC	0.00	-0.09	0.10	0.16	0.04	0.32	<u>0.50</u>	0.17	0.33
DO	-0.18	-0.01	-0.23	0.36	<u>0.53</u>	-0.08	-0.04	<u>0.54</u>	-0.10
pH	0.06	0.17	-0.16	0.35	0.42	0.03	-0.02	0.45	-0.10
chl-a	0.23	0.25	0.04	<u>0.55</u>	<u>0.69</u>	0.11	0.08	<u>0.63</u>	0.06
Sites B									
WT	0.00	-0.04	0.30	-0.16	-0.32	<u>0.49</u>	0.41	-0.04	0.40
SD	-0.42	-0.43	0.09	<u>-0.77</u>	<u>-0.88</u>	0.11	0.14	<u>-0.75</u>	0.25
EC	-0.31	-0.25	-0.06	-0.11	-0.10	0.12	0.06	-0.22	0.11
DO	0.33	0.27	-0.14	<u>0.59</u>	<u>0.72</u>	-0.30	-0.21	<u>0.52</u>	-0.35
pH	0.12	0.10	-0.20	0.45	0.34	-0.15	-0.19	0.36	-0.20
chl-a	0.20	0.06	0.04	<u>0.75</u>	<u>0.77</u>	0.04	0.11	<u>0.63</u>	-0.01

(Sites A – upstream the lock, Sites B – downstream the lock)

<math>p<0.05</math>; statistical significances are underlined

The hierarchical clustering analysis showed that sites located upstream of locks (1A, 2A, 3A) are more similar to each other than to other sites, as are the sites located downstream of locks (1B, 2B, 3B). The first dendrogram grouped the sites based on similar environmental parameters (physico-chemical data). The second dendrogram grouped the sites based on similar biological parameters (zooplankton data) (Figure 2A and Figure 2B).

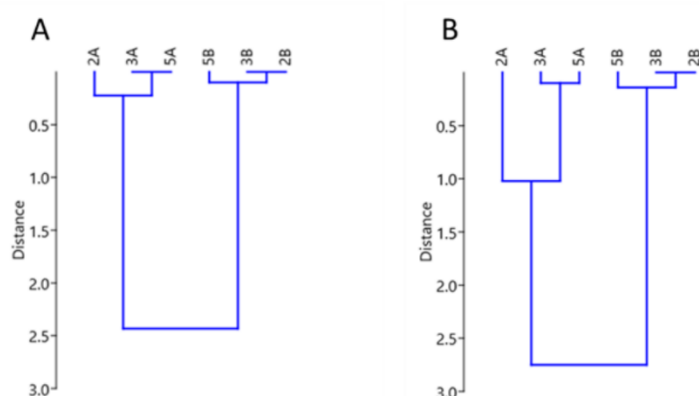


Figure 2A and 2B. Tree diagrams of cluster analysis of study sites based on environmental parameters (A) obtained using Euclidean similarity index and biological parameters (B) (zooplankton data) obtained by the Bray–Curtis similarity index.

RDA analysis described the relationship between zooplankton species composition and environmental parameters in the Bydgoszcz Canal and the Noteć Canal upstream and

downstream of locks. The length of environmental parameters (vectors) indicated their significance alongside the corresponding axes. Statistical significances described below are confirmed by results of RDA using the Pearson test. Correlations and RDA were calculated based on the original dataset. At upstream sites, the results of the ordination showed that the eigenvalues of the first ($\lambda_{RDA1}=0.357$) and second ($\lambda_{RDA2}=0.085$) RDA axes accounted for 44.2% of the variation in the environmental data. The longest vector Secchi disk visibility showed highest negative correlation with total zooplankton density ($r=-0.72$, $p<0.05$), total rotifers density ($r=-0.83$, $p<0.05$) and their biomass ($r=-0.84$, $p<0.05$). Conversely, dissolved oxygen showed a positive relationship with total rotifer density ($r=0.53$, $p<0.05$) and total rotifer biomass ($r=0.54$, $p<0.05$). Chlorophyll concentration had also a positive relationship with density ($r=0.69$, $p<0.05$), total rotifer biomass ($r=0.63$, $p<0.05$) and total zooplankton density ($r=0.55$, $p<0.05$). Conductivity exhibited a positive relationship with zooplankton biomass ($r=0.50$, $p<0.05$) (Table 4). The vectors for pH and water temperature were short, which means that these parameters are of less importance and statistically not significant in correlation with zooplankton species (Figure 3).

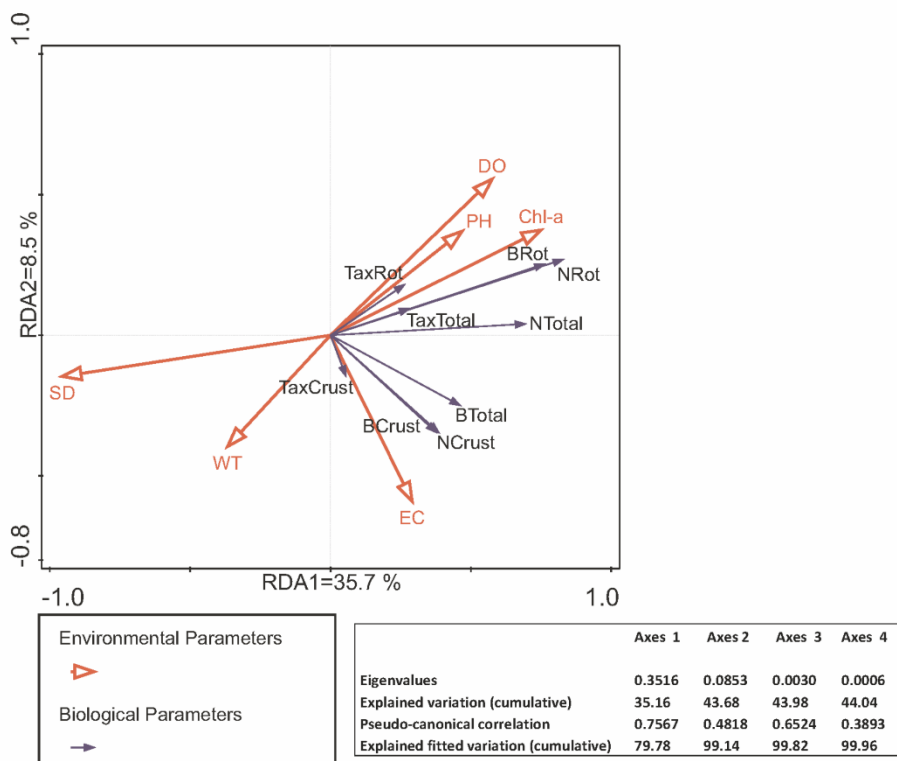


Figure 3. Results of Redundancy analysis (RDA) performed on zooplankton and environmental data during growing season in the Bydgoszcz Canal and the Noteć Canal

1 upstream of locks using forward selection of variables ($p < 0.05$). Triplot of significant
2 environmental variables (WT – Water temperature, SD – Secchi disk visibility, EC –
3 conductivity, DO – oxygen concentration, pH, Chl-a – chlorophyll-a), number of zooplankton
4 species (Tax total, Tax Rot, Tax Crust), zooplankton density (N total, N Rot, N Crust) and
5 zooplankton biomass (B total, B Rot, B Crust)
6
7
8
9

10 At sites downstream of locks, the results of the ordination showed that the eigenvalues of the
11 first ($\lambda_{RDA1} = 0.345$) and second ($\lambda_{RDA2} = 0.181$) RDA axes accounted for 52.6% of the
12 variation in the environmental data. The longest vector Secchi disk visibility showed highest
13 negative correlation with rotifers density ($r = -0.88$, $p < 0.05$) their biomass ($r = -0.75$, $p < 0.05$) and
14 zooplankton density ($r = -0.77$, $p < 0.05$). On the other hand, water temperature was positively
15 correlated with crustacean density ($r = 0.49$, $p < 0.05$). Dissolved oxygen showed a positive
16 relationship with total zooplankton density ($r = 0.59$, $p < 0.05$), rotifer density ($r = 0.72$, $p < 0.05$)
17 and rotifer biomass ($r = 0.52$, $p < 0.05$). Chlorophyll concentration also had a positive relationship
18 with total zooplankton density ($r = 0.75$, $p < 0.05$), rotifer density ($r = 0.77$, $p < 0.05$) and rotifer
19 biomass ($r = 0.63$, $p < 0.05$) (Table 4). The vectors for pH and conductivity were short, which
20 means that these parameters are of less importance and statistically not significant in correlation
21 with zooplankton species (Figure 4).
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

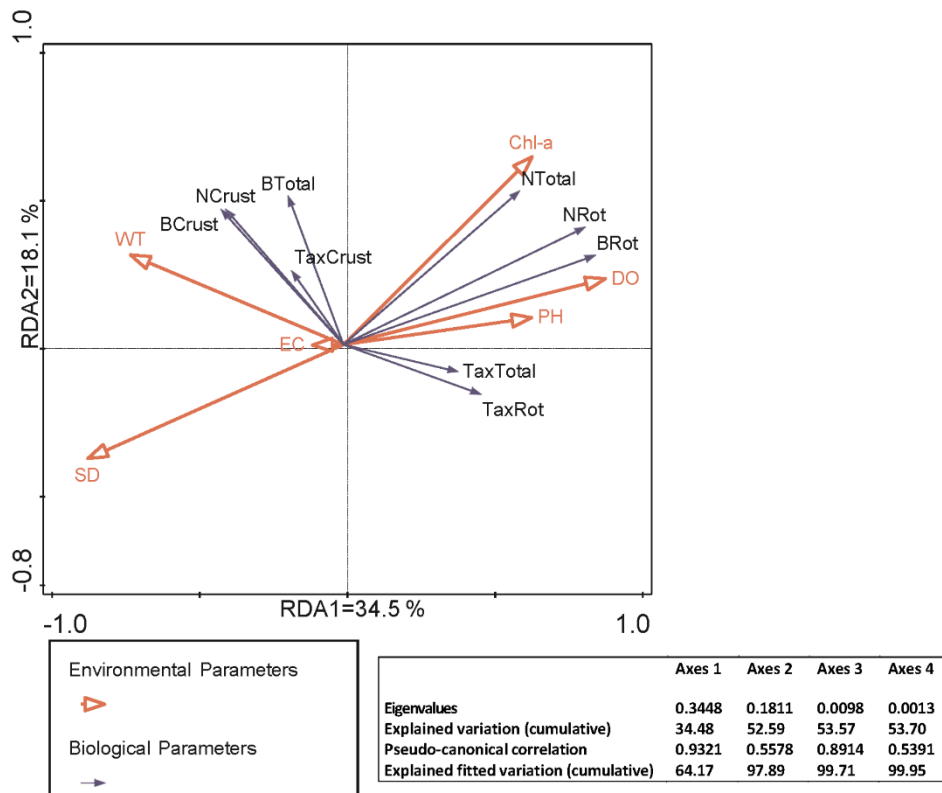


Figure 4. Results of Redundancy analysis (RDA) performed on zooplankton and environmental data during vegetation season in the Bydgoszcz Canal and the Noteć Canal downstream of locks using forward selection of variables ($p < 0.05$). Triplot of significant environmental variables (WT – Water temperature, SD – Secchi disk visibility, EC – conductivity, DO – oxygen concentration, pH, Chl-a – chlorophyll-a), number of zooplankton species (Tax total, Tax Rot, Tax Crust), zooplankton density (N total, N Rot, N Crust) and zooplankton biomass (B total, B Rot, B Crust)

Discussion

Zooplankton was studied in the Bydgoszcz Canal and the Noteć Canal upstream and downstream of the selected locks in 2021 and 2022. The hierarchical cluster analysis showed differences in environmental and biological conditions between habitats located upstream and downstream of locks. According to the results, the environmental (Figure 2A) and biological

1 conditions (Figure 2B) at sites upstream of locks in Bydgoszcz Canal and Noteć Canal were
2 distinct from the sites downstream of the locks (Figure 2A and Figure 2B).

3 RDA analysis identified that environmental variables were closely related to changes
4 in the zooplankton community structure and niche differentiation. Chlorophyll concentration
5 and water temperature were the main limiting factors and played important roles in shaping
6 the zooplankton community structure (Figure 3 and Figure 4).

7 The average values of environmental parameters were similar upstream and
8 downstream of locks on the studied canals. However, our study showed that chlorophyll
9 concentration was higher downstream of the lock at site 1 in the Bydgoszcz Canal than
10 upstream (Table 1). The increase in chlorophyll concentration indicated an increase in
11 zooplankton density (rotifers and crustaceans) at this site. According to Sommer et al. (2004)
12 the high chlorophyll concentration suggested greater food availability for zooplankton
13 development in the floodplain lake. The level of chlorophyll (phytoplankton) increased during
14 high flow, particularly in shallow water habitats (such as in the Bydgoszcz Canal downstream
15 of locks).

16 In our study, we noticed that zooplankton diversity was higher at sites upstream of locks
17 than downstream (Table 2 and Table 3).

18 During the study, rotifers dominated over crustaceans in density and diversity. The
19 number of rotifer species upstream slightly prevailed of the number downstream (Table 2 and
20 Table 3). Rotifers represented 80% of all zooplankton species. The most dominant were *K.*
21 *cochlearis*, *K. quadrata* and *K. tecta*.

22 Crustaceans represented 20% of all zooplankton species. The most dominant among
23 crustaceans were Cladocera – *B. longirostris*, *C. pulchella*, *Ch. sphaericus* and nauplii
24 (copepod larval forms) (Table 2). Similar zooplankton species have been observed in a slow-
25 flowing section of the lower Oder river (Czerniawski et al., 2013) and in man-made ditches
26 (Czerniawski and Sługocki, 2017). In those studies, the authors confirmed that slow water flow
27 favored the development of zooplankton communities.

28 Kuczyńska-Kippen et al. (2021) suggested that high rotifer density was associated with
29 the presence of submerged macrophytes. Thus, the distribution and abundance of rotifers are
30 influenced by the presence of submerged macrophytes and the flow velocity of the water.
31 Therefore, the sections upstream of locks (low flow velocity and high macrophyte vegetation)
32 favored high density and diversity of rotifers. Several authors have reported that rotifers are
33

1 less susceptible than crustaceans to changes in environmental and hydrological conditions
2 (Marneffe et al., 1996; Demetraki-Paleolog, 2004; Špoljar et al., 2012).

3
4 The zooplankton biomass, including rotifer biomass, was also higher upstream of locks
5 than downstream (site 2 and site 3). This happened because the zooplankton density increased
6 at these sites. According to some authors, the zooplankton biomass in turbulent is much lower
7 waters (e.g., downstream of locks) than in calm waters (e.g., upstream of locks) (Baranayi et
8 al., 2002; Dickerson et al., 2010; Czerniawski and Domagała, 2012; Zhou et al., 2016).
9 Therefore, the zooplankton community is probably shaped primarily by intensity of water
10 movement. Only site 1 showed higher total zooplankton density, rotifer density, crustacean
11 density, total zooplankton biomass and crustacean biomass downstream of locks than
12 upstream.
13
14
15
16
17
18
19

20 Presumably, this difference is the result of internal loading (organic matter and
21 nutrients) after the re-suspension of bottom sediments due to increased water movement inside
22 the lock. Similar changes have been reported by Jeppesen et al. (2014). The sediment
23 deposition downstream of the lock creates a more nutrient-rich environment, which can support
24 increased primary production, including algal growth (higher chlorophyll concentration)
25 (Cottingham et al., 1997; Chaparro et al., 2011). The increased availability of food resources
26 can contribute to higher densities and biomass of rotifers and crustaceans (Kolarova &
27 Napiórkowski, 2022).
28
29
30
31
32
33
34

35 RDA analysis indicated that the density and biomass of rotifers changed similarly to
36 chlorophyll at sites upstream of locks (Figure 3). Our results support the hypothesis that rotifer
37 communities are defined by bottom-up effects linked to food supply, such as small
38 phytoplankton (chlorophyll – probably Cryptophyta) (Yoshida et al., 2003; Felpeto et al., 2013;
39 Dembowska, 2021; Wang et al., 2022). Shayestehfar et al. (2008) pointed out that rotifer
40 density and distribution depend on the variety of ecological and physicochemical factors such
41 as food availability, but also dissolved oxygen and pH. All these factors are essential to
42 determine the variations in rotifer density (Li et al., 2014; Wang et al., 2016; Banerjee et al.,
43 2019; Kim et al., 2022).
44
45
46
47
48
49
50

51 According to RDA analysis, the main factor affecting the density and biomass of
52 crustacean zooplankton, as well as total zooplankton biomass, at sites downstream of locks was
53 water temperature (Figure 4). This relationship was observed in the triplot of significant
54 environmental variables and was confirmed by Pearson's simple correlation r coefficient.
55
56
57
58
59
60
61
62
63
64
65

1 Similarly, a study by Wei et al. (2017) showed that crustacean zooplankton was
2 positively correlated with water temperature in a large river–lake system. Moore et al. (1996)
3 suggested that temperature is an important compositional factor for crustaceans because
4 temperature controls feeding, respiration, egg production velocity and other metabolic
5 processes. All studied sites were similar in terms of abundance and biomass, except for site 1
6 downstream of the lock. The effects that water movement inside locks have on environmental
7 conditions and zooplankton development need further study.
8
9
10
11
12
13
14
15

16 **Conclusions**

17
18
19
20 In the studied canals, zooplankton was shaped by different environmental and hydrological
21 conditions.
22

- 23 • Density and biomass of rotifers changed similarly to chlorophyll at sites upstream of locks.
- 24 • Zooplankton diversity also increased upstream of locks.
- 25 • Low flow velocity and macrophytes as habitat favored high diversity and density of
26 zooplankton.
- 27 • Water temperature stimulated the growth of crustaceans (density and biomass) and the total
28 zooplankton biomass at sites downstream of locks.
- 29 • Only site 1 zooplankton density (rotifers and crustaceans), zooplankton biomass and
30 crustacean biomass to be higher downstream of the lock than upstream. This is likely the
31 result of internal loading (organic matter and nutrients) after the re-suspension of bottom
32 sediments due to increased water movement inside the lock.
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

References

- 1
2
3
4 Appel, D. S., Gerrish, G. A., Fisher, E. J., Fritts, M. W., 2020. Zooplankton sampling in large
5 riverine systems: A gear comparison. *River Research and Applications*, 36(1), 102-114.
6 doi:10.1002/rra.3539
7
8
9
10 Ball, E. E., Smith, D. E., Anderson, E. J., Skufca, J. D., Twiss, M. R., 2018. Water velocity
11 modeling can delineate nearshore and main channel plankton environments in a large
12 river. *Hydrobiologia*, 815, 125-140. doi:10.1007/s10750-018-3556-5
13
14
15
16 Banerjee, A., Chakrabarty, M., Rakshit, N., Bhowmick, A. R., Ray, S., 2019. Environmental
17 factors as indicators of dissolved oxygen concentration and zooplankton abundance: Deep
18 learning versus traditional regression approach. *Ecological Indicators*, 100, 99-117.
19 doi:10.1016/j.ecolind.2018.09.051
20
21
22
23
24
25 Baranyi, C., Hein, T., Holarek, C., Keckeis, S., Schiemer, F., 2002. Zooplankton biomass and
26 community structure in a Danube River floodplain system: effects of
27 hydrology. *Freshwater Biology*, 47(3), 473-482.
28
29
30
31 Błędzki, L.A., Rybak J.I., 2016. *Freshwater Crustacean Zooplankton of Europe. Cladocera and*
32 *Copepoda (Calanoida, Cyclopoida) Key to species identification. Springer International*
33 *Publishing Switzerland*, doi:10.1007/978-3-319-29871-9_7
34
35
36
37
38 Chaparro, G., Fontanarrosa, M. S., Schiaffino, M. R., de Tezanos Pinto, P., O'Farrell, I., 2014.
39 Seasonal-dependence in the responses of biological communities to flood pulses in warm
40 temperate floodplain lakes: implications for the "alternative stable states" model. *Aquatic*
41 *sciences*, 76, 579-594. doi:10.1007/s00027-014-0356-5
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

- 1 Czerniawski, R., Domagała, J., 2012. Potamozooplankton communities in three different
2 outlets from mesotrophic lakes located in lake-river system. *Oceanological and*
3 *Hydrobiological Studies*, 41(1), 46–56. doi:10.2478/s13545-012-0006-2
4
5
6 Czerniawski, R., Pilecka-Rapacz, M., Domagała, J., 2013. Zooplankton communities of inter-
7 connected sections of lower River Oder (NW Poland). *Open Life Sciences*, 8(1), 18-29.
8 doi:10.2478/s11535-012-0110-8
9
10
11
12 Czerniawski, R., Sługocki, Ł., 2017. Analysis of zooplankton assemblages from man-made
13 ditches in relation to current velocity. *Oceanological and Hydrobiological Studies*, 46(2),
14 199-211. doi:10.1515/ohs-2017-0020
15
16
17
18
19 Dembowska, E.A., 2021. The Use of Phytoplankton in the Assessment of Water Quality in
20 the Lower Section of Poland’s Largest River. *Water*, 13, 3471. doi:10.3390/w13233471
21
22
23 Demetraki-Paleolog, A., 2004. Planktonic rotifers diversity in selected rivers of the Vistula,
24 Wieprz and San drainage-basins. *Teka Kom. Ochr. Kszt. Środ. Przyr. – OL PAN*, 1, 44-
25 50.
26
27
28
29
30 Dickerson, K. D., Medley, K. A., Havel, J. E., 2010. Spatial variation in zooplankton
31 community structure is related to hydrologic flow units in the Missouri River, USA. *River*
32 *Research and Applications*, 26(5), 605-618. doi:10.1002/rra.1268
33
34
35
36
37 Dorotovičová, C., 2013. Man-made canals as a hotspot of aquatic macrophyte biodiversity in
38 Slovakia. *Limnologica*, 43(4), 277-287. doi:10.1016/j.limno.2012.12.002
39
40
41
42 Einsle, U., 1996. Copepoda: Cyclopoida: genera Cyclops, Megacyclops, Acanthocyclops. In
43 *Guides to the identification of the microinvertebrates of the continental waters of the*
44 *world*; Dumont, H.J. Ed. 10. 83 pp., Amsterdam SPB Academic Publishing by. Nederland.
45
46
47
48 Felpeto, A. B., Hairston, Jr, Nelson G., 2013. Indirect bottom- up control of consumer-
49 resource dynamics: resource- driven algal quality alters grazer numerical
50 response. *Limnology and oceanography*, 58(3), 827-838. doi:10.4319/lo.2013.58.3.0827
51
52
53
54
55 Furst, D. J., Aldridge, K. T., Shiel, R. J., Ganf, G. G., Mills, S., Brookes, J. D., 2014. Floodplain
56 connectivity facilitates significant export of zooplankton to the main River Murray channel
57 during a flood event. *Inland Waters*, 4(4), 413-424. doi:10.5268/IW-4.4.696
58
59
60
61
62
63
64
65

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- Gianuca, A. T., Pantel, J. H., De Meester, L., 2016. Disentangling the effect of body size and phylogenetic distances on zooplankton top-down control of algae. *Proceedings of the Royal Society B: Biological Sciences*, 283(1828), 20160487. doi:10.1098/rspb.2016.0487
- Gorączko, M., 2015. Natężenie ruchu żeglugowego na bydgoskim odcinku drogi wodnej E-70-stan obecny i perspektywy rozwoju. *Geography and Tourism*, 1(3), 33-38.
- Hammer, Ø., Harper, D.A.T., Ryan, P.D., 2001. PAST: Paleontological Statistic software package for education and data analysis. *Paleontologia Electronica*, 4 (1), 178.
- Harris, R.P., Wiebe, P.H., Lenz, J., Skjoldal, H.R., Huntley, M., 2000. ICES Zooplankton methodology manual. Academic Press Elsevier Ltd, 684 pp. doi:10.1016/B978-0-12-327645-2.X5000-2
- Havel, J. E., Medley, K. A., Dickerson, K. D., Angradi, T. R., Bolgrien, D. W., Bukaveckas, P. A., Jicha, T. M., 2009. Effect of main-stem dams on zooplankton communities of the Missouri River (USA). *Hydrobiologia*, 628, 121-135. doi:10.1007/s10750-009-9750-8
- Jeelani, M., Kaur, H., Kumar, R., 2008. Impact of Climate Warming on the Biodiversity of Freshwater Ecosystem of Kashmir, India. In *Proceedings of Taal2007: The 12th World Lake Conference (Vol. 1103, p. 1109)*.
- Jeppesen, E., Meerhoff, M., Davidson, T. A., Trolle, D., Sondergaard, M., Lauridsen, T. L., Beklioğlu M., Brucet, S., Volta, P., Gonzáles-Bengonzoni, I., Nielsen, A., 2014. Climate change impacts on lakes: an integrated ecological perspective based on a multi-faceted approach, with special focus on shallow lakes. *Journal of Limnology*, 73(s1): 88-111 doi:10.4081/jlimnol.2014.844
- Kelly, L. A., Hassall, C., 2018. The spatial ecology of phytoplankton blooms in UK canals. *Inland Waters*, 8(4), 422-433. doi:10.1080/20442041.2018.1482152
- Kelly, N. E., Wantola, K., Weisz, E., Yan, N. D., 2013. Recreational boats as a vector of secondary spread for aquatic invasive species and native crustacean zooplankton. *Biological Invasions*, 15, 509-519. doi:10.1007/s10530-012-0303-0
- Kim, J., Mandrak, N. E., 2016. Assessing the potential movement of invasive fishes through the Welland Canal. *Journal of Great Lakes Research*, 42(5), 1102-1108. doi:10.1016/j.jglr.2016.07.009

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- Kim, S. K., Yun, J. H., Joo, G. J., Choi, J. Y., 2022. Hydrological Characteristics and Trophic Status as Dominant Drivers of Rotifer Community Composition in Artificially Created Riverine Wetlands. *Animals*, 12(4), 461. doi:10.3390/ani12040461
- Kolarova, N., Napiórkowski, P., 2022. How Do Specific Environmental Conditions in Canals Affect the Structure and Variability of the Zooplankton Community?. *Water*, 14(6), 979. doi:10.3390/w14060979
- Kuczyńska-Kippen, N., Špoljar, M., Mleczek, M., Zhang, C., 2021. Elodeids, but not helophytes, increase community diversity and reduce trophic state: Case study with rotifer indices in field ponds. *Ecological Indicators*, 128, 107829. doi:10.1016/j.ecolind.2021.107829
- Li, X., Yu, H., Ma, C., 2014. Zooplankton community structure in relation to environmental factors and ecological assessment of water quality in the Harbin Section of the Songhua River. *Chinese Journal of Oceanology and Limnology*, 32(6), 1344-1351. doi:10.1007/s00343-014-3303-3
- Majagi, S., Naik, J., Chitra, J., 2019. Seasonal Investigation on the Zooplankton Diversity and Distribution in Relation to Water Quality at Chikklingdalli Dam, Karnataka. *Int. J. Res. Anal. Rev*, 1, 754-767.
- Manickam, N., Bhavan, P. S., Santhanam, P., Bhuvanewari, R., Muralisankar, T., Srinivasan, V., Asaikkutti A., Rajkumar G., Udayasuriyan R., Karthik, M., 2018. Impact of seasonal changes in zooplankton biodiversity in Ukkadam Lake, Coimbatore, Tamil Nadu, India, and potential future implications of climate change. *The Journal of Basic and Applied Zoology*, 79, 1-10. doi:10.1186/s41936-018-0029-3
- Marneffe, Y., Descy, J. P., Thomé, J. P., 1996. The zooplankton of the lower river Meuse, Belgium: seasonal changes and impact of industrial and municipal discharges. *Hydrobiologia*, 319, 1-13. doi:10.1007/BF00020966
- McCabe, G. T., Hinton, S. A., Emmett, R. L., 1998. Benthic invertebrates and sediment characteristics in a shallow navigation channel of the lower Columbia River, before and after dredging. *Northwest science*, 72(2), 116-126.

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- McCauley, E., 1984. The estimation of the abundance and biomass of zooplankton in samples. A manual on methods for the assessment of secondary productivity in fresh waters.
- Medeiros, E. S., Arthington, A. H., 2008. The importance of zooplankton in the diets of three native fish species in floodplain waterholes of a dryland river, the Macintyre River, Australia. *Hydrobiologia*, 614, 19-31. doi:10.1007/s10750-008-9533-7
- Moore, M. V., Folt, C. F., Stemberger, R. S., 1996. Consequences of elevated temperatures for zooplankton assemblages in temperate lakes. *Archiv für Hydrobiologie*, 289-319. doi:10.1127/archiv-hydrobiol/135/1996/289
- Muszyńska-Jeleszyńska, D., Marciniak, Ż., 2016. Kanał Bydgoski i kanał Finow-Jedna historia. Dwa kanały. *Wspólna przyszłość. Gospodarka Wodna*, (12), 430-433.
- Napiórkowski, P., Bąkowska, M., Mrozińska, N., Szymańska, M., Kolarova, N., Obolewski, K., 2019. The effect of hydrological connectivity on the zooplankton structure in floodplain lakes of a regulated large river (the lower Vistula, Poland). *Water*, 11 (9), 1924. doi:10.3390/w11091924.
- Nguyen, T., Roddick, F.A., Fan, L., 2015. Impact of green algae on the measurement of 18 *Microcystis aeruginosa* populations in lagoon-treated wastewater with an algae online 19 analyser. *Environmental Technology*, 36(5), 556–565. doi:10.1080/09593330.2014.20953212
- Ning, N.S., Nielsen, D.L., Hillman, T.J., Suter, P.J., 2010. The influence of planktivorous fish on zooplankton communities in riverine slackwaters. *Freshwater Biology*, 55 (2), 360-374. doi:10.1111/j.1365-2427.2009.02283.x.
- Onikura, N., 2015. Site selection for habitat conservation/restoration of threatened freshwater fishes in artificial channels of northern Kyushu Island, Japan. *Ichthyological Research*, 62, 197-206. doi:10.1007/s10228-014-0427-6
- Pagnucco, K. S., Maynard, G. A., Fera, S. A., Yan, N. D., Nalepa, T. F., Ricciardi, A., 2015. The future of species invasions in the Great Lakes-St. Lawrence River basin. *Journal of Great Lakes Research*, 41, 96-107. doi:10.1016/j.jglr.2014.11.004

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- Radwan, S., Bielańska-Grajner I., Ejsmont-Karabin, J., 2004. Rotifers (Rotifera, Monogononta), Freshwater Fauna of Poland University of Lodz Press: Łódź, Poland, pp. 579.
- Rao, C. R., 1964. The use and interpretation of principal component analysis in applied research. *Sankhyā: The Indian Journal of Statistics, Series A*, 329-358.
- Roberts, J. J., Rahel, F. J., 2008. Irrigation canals as sink habitat for trout and other fishes in a Wyoming drainage. *Transactions of the American Fisheries Society*, 137(4), 951-961. doi:10.1577/T07-058.1
- Shayestehfar, A., Soleimani, M., Mousavi, S. N., Shirazi, F., 2008. Ecological study of rotifers from Kor river, Fars, Iran. *Journal of Environmental Biology*, 29(5), 715-720.
- Segovia, P., Rajaoarisoa, L., Nejjari, F., Duviella, E., Puig, V., 2019. Model predictive control and moving horizon estimation for water level regulation in inland waterways. *Journal of Process Control*, 76, 1-14. doi:10.1016/j.jprocont.2018.12.017
- Sommer, T. R., Harrell, W. C., Solger, A. M., Tom, B., Kimmerer, W., 2004. Effects of flow variation on channel and floodplain biota and habitats of the Sacramento River, California, USA. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 14(3), 247-261. doi:10.1002/aqc.620
- Symons, C. C., Arnott, S. E., 2013. Regional zooplankton dispersal provides spatial insurance for ecosystem function. *Global Change Biology*, 19(5), 1610-1619. doi:10.1111/gcb.12122
- Špoljar, M., Dražina, T., Šargač, J., Borojević, K. K., Žutinić, P., 2012. Submerged macrophytes as a habitat for zooplankton development in two reservoirs of a flow-through system (Papuk Nature Park, Croatia). In *Annales de Limnologie-International Journal of Limnology*, (Vol. 48, No. 2, pp. 161-175). EDP Sciences. doi:10.1051/limn/2012005
- Tarkowska-Kukuryk, M., Grzywna, A., 2022. Macrophyte communities as indicators of the ecological status of drainage canals and regulated rivers (Eastern Poland). *Environmental Monitoring and Assessment*, 194(3), 210. doi:10.1007/s10661-022-09777-0

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- ter Braak, C.J.F., Šmilauer, P., 2002. CANOCO reference manual and CanoDraw for windows 33 user's guide: Software for canonical community ordination (version 4.5). Microcomputer 34 Power, Ithaca, New York.
- van den Wollenberg, A. L., 1977. Redundancy analysis an alternative for canonical correlation analysis. *Psychometrika*, 42(2), 207-219.
- Venkatesharaju, K., Ravikumar, P., Somashekar, R. K., Prakash, K. L., 2010. Physico-chemical and bacteriological investigation on the river Cauvery of Kollegal stretch in Karnataka. *Kathmandu University Journal of Science, Engineering and Technology*, 6(1), 50-59.
- Wallace R.L., Snell T.W., Ricci C., Nogrady, T., 1993. Rotifera [Ed. Thomas Nogrady] Volume 1: Biology, Ecology and Systematics. In: *Guides to the Identification of the Microinvertebrates of the Continental Waters of the World* [Ed. Henri J. Dumont], SPB Academic Publishing bv, The Hague.
- Wang, W., Shi, K., Zhang, Y., Li, N., Sun, X., Zhang, D., Zhang, Y., Qin B., Zhu, G., 2022. A ground-based remote sensing system for high-frequency and real-time monitoring of phytoplankton blooms. *Journal of Hazardous Materials*, 439, 129623. doi:10.1016/j.jhazmat.2022.129623
- Wang, C., Wang, L., Deng, D., Zhou, Z., 2016. Temporal and spatial variations in rotifer correlations with environmental factors in Shengjin Lake, China. *Environmental Science and Pollution Research*, 23, 8076-8084. doi:10.1007/s11356-015-6009-y
- Weber, A., Garcia, X. F., Wolter, C., 2017. Habitat rehabilitation in urban waterways: the ecological potential of bank protection structures for benthic invertebrates. *Urban Ecosystems*, 20, 759-773. doi:10.1007/s11252-017-0647-4
- Wei, W., Chen, R., Wang, L., Fu, L., 2017. Spatial distribution of crustacean zooplankton in a large river-connected lake related to trophic status and fish. *Journal of limnology*, 76(3). doi:10.4081/jlimnol.2017.1622
- Yoshida, T., Urabe, J., Elser, J. J., 2003. Assessment of 'top-down' and 'bottom-up' forces as determinants of rotifer distribution among lakes in Ontario, Canada. *Ecological Research*, 18, 639-650. doi:10.1111/j.1440-1703.2003.00596.x

1
2 Zhao, K., Wang, L., Riseng, C., Wehrly, K., Pan, Y., Song, K., Da, I., Pang, W., You, Q., Tian,
3
4 H., Liu, S., Wang, Q., 2018. Factors determining zooplankton assemblage difference
5
6 among a man-made lake, connecting canals, and the water-origin river. *Ecological*
7
8 *Indicators*, 84, 488 doi:10.1016/j.ecolind.2017.07.052
9

10
11 Zhou, J., Han, X., Qin, B., Casenave, C., Yang, G., 2016. Response of zooplankton community
12
13 to turbulence in large, shallow Lake Taihu: a mesocosm experiment. *Fundamental and*
14
15 *Applied Limnology*, 187(4), 315-324.
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Author Agreement Statement

We the undersigned declare that this manuscript is original, has not been published before and is not currently being considered for publication elsewhere.

We confirm that the manuscript has been read and approved by all named authors and that there are no other persons who satisfied the criteria for authorship but are not listed. We further confirm that the order of authors listed in the manuscript has been approved by all of us.

We understand that the Corresponding Author is the sole contact for the Editorial process. He/she is responsible for communicating with the other authors about progress, submissions of revisions and final approval of proofs

Signed by all authors as follows:

Kolarawal
Rajiv

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Pawł Napiórkowski
miron Kotarona